# Appendix AB-O, Tributary Habitat Restoration **Attachment O.3 Sacramento River Weighted Usable Area Analysis**

# **O.3.1 Model Overview**

Weighted usable area (WUA) analysis is a method for estimating the availability of suitable habitat in rivers, streams, and floodplains under different flow conditions (Bovee et al. 1998). It has been used primarily for estimating spawning and rearing habitat of fish species. WUA is computed as the surface area of physical habitat available for spawning or rearing, weighted by its suitability. Habitat suitability is determined from field studies of the distributions of redds or rearing juveniles with respect to flow velocities, depths, and substrate or cover in the river or floodplain (Bovee et al. 1998). These data are used in hydraulic and habitat model simulations (e.g., PHABSIM or RIVER2D) that estimate the availability of suitable habitat in a portion of the river at a given flow. WUA curves showing suitable habitat availability versus flow are generated from the simulations. These curves facilitate evaluating how different flow regimes affect spawning and rearing habitat of important fish species.

# **O.3.2 Model Development**

#### **O.3.2.1 Methods**

For this analysis, spawning and rearing WUA was estimated for winter-run, spring-run, fall-run, and late fall–run Chinook salmon and California Central Valley steelhead in the Sacramento River. Spawning and rearing WUA were estimated for the BA and EIS modeled scenarios from CalSim 3 flow data for each month of the 93-year period of record.

#### *O.3.2.1.1 Sacramento River Spawning WUA*

The WUA curves used for Chinook salmon and steelhead spawning habitat in the Sacramento River were obtained from three U.S. Fish and Wildlife Service reports (U.S. Fish and Wildlife Service 2003a, 2005a, 2006). Modeling assumptions used to derive spawning WUA curves include that the suitability of physical habitat for salmon and steelhead spawning is largely a function of substrate particle size, water depth, and flow velocity. The race- or species-specific suitability of the habitat with respect to these variables is determined by cataloguing conditions at active redds and is used to develop habitat suitability criteria (HSC) for each race or species of fish. Hydraulic modeling is then used to estimate the amount of habitat available for different HSC levels at different river flows, and the results are combined to develop spawning habitat WUA curves (Bovee et al. 1998). The WUA curves and tables are used to look up the amount of spawning WUA available at different flows during the spawning periods of the race or species.

U.S. Fish and Wildlife Service (2003a) provides WUA curves and tables for spawning winterrun, fall-run, and late fall-run Chinook salmon and steelhead for three segments of the Sacramento River encompassing the reach from Keswick Dam to Battle Creek (Figure O.3-1). U.S. Fish and Wildlife Service (2005a) provides WUA curves and tables for spawning fall-run in an additional downstream segment (Battle Creek to the former location of the Red Bluff Diversion Dam  $[RBDD]$ <sup>1</sup>) because spawning for fall-run occurs further downstream than it does for the other races of salmon (Figure O.3-1). The PHABSIM hydraulic model was used for these studies. All WUA tables were updated in 2006 using the more recently developed RIVER2D model (U.S. Fish and Wildlife Service 2006). No spawning WUA curves were developed for spring-run Chinook salmon, so the fall-run curves were used to quantify spring-run spawning habitat. The basis and potential uncertainties of this substitution are discussed below in Section O.3.2.2, *Assumptions/Uncertainties*. Although fall-run spawning WUA curves were used as surrogates for spring-run spawning, CalSim 3 flows for the months of spring-run spawning, not those of fall-run spawning, were used to compute the spring-run WUA results. Also, the HSC used to develop the steelhead WUA curves for Sacramento River spawning were obtained from investigations of steelhead redds in the American River (U.S. Fish and Wildlife Service 2003b). The need for and uncertainty of this substitution are also discussed in Section O.3.2.2, *Assumptions/Uncertainties*.

Figure O.3-2 through Figure O.3-5 show the flow versus spawning WUA results for winter-run, fall-run, late fall-run, and steelhead in the three upstream river segments (Segment  $6 =$ Keswick to Anderson-Cottonwood Irrigation District [ACID] Dam, Segment 5 = ACID Dam to Cow Creek, and Segment 4 = Cow Creek to Battle Creek) as provided by U.S. Fish and Wildlife Service (2003a). Figure O.3-6 shows spawning WUA results for fall-run in the more downstream segment (Segment 3 = Battle Creek to RBDD (U.S. Fish and Wildlife Service 2005a). Note that for Segment 6, separate WUA curves were developed for periods when the ACID Dam boards were installed (April through October) and for when the boards were out because installation of the boards affects water depths and flow velocities for some of the sampling transects used to develop the curves.

Several tributaries enter the Sacramento River between Keswick Dam and Battle Creek, resulting in differences in flow among the river segments. For the U.S. Fish and Wildlife Service studies, Sacramento River flows were measured directly at the sampling transects and were estimated as the sum of Keswick Dam flow releases and tributary gauge readings upstream of the transects. For the WUA analyses used in this analysis, the segment flows were estimated using Sacramento River CalSim 3 flows at Keswick Dam and the Clear Creek, Cow Creek, and Battle Creek confluences. Keswick Dam flows were used for Segment 6 and for Segment 5 upstream of the Clear Creek confluence. Flows at Clear Creek were used for Segment 5 downstream of the confluence. Flows at Cow Creek were used for Segment 4 and flows at Battle Creek were used for Segment 3. For Segment 6, the WUA curves for the months when the ACID Dam boards are installed (April through October) were used with the Keswick Dam flows for those months and

 $1$  For simplicity, this location is referred to as the Red Bluff Diversion Dam (RBDD) in this document despite dam decommissioning in 2013.

the WUA curves for the months when the ACID Dam boards are out were used with the flows for the rest of the year.

Mean spawning WUA for this analysis were examined for the principal months of winter-run, spring-run, fall-run, late fall-run and steelhead spawning periods (Table O.3-1) under each water year type and all water year types combined. Total spawning WUA for all months combined was computed by weighting the monthly results by monthly weighting factors (Table O.3-1). For winter-run and late fall-run, these weighting factors were estimated from the mean proportions of redds counted each months in the aerial redd surveys conducted by California Department of Fish and Wildlife during 2006 through 2021 (CDFW unpublished data). Information from Williams (2006) was also used in estimating the late fall-run spawning months. For spring-run and steelhead the weighting factors were derived from information on life-history timings of listed anadromous salmonids of the Central Valley in Appendix AB-C, and for fall-run the weighting factors were estimated from information in Moyle et al. 2017.



Table O.3-1. Monthly Weighting Factors for Sacramento River Winter-run, Spring-run, Fall-run, Late fall-run, and Steelhead Spawning.



ACID = Anderson-Cottonwood Irrigation District Source: U.S. Fish and Wildlife Service 2003a.

Figure O.3-1. Segments 2–6 of the Sacramento River Used in U.S. Fish and Wildlife Service Studies to Determine Spawning and Rearing WUA (flows in the figure are the average flows at the upstream boundary of each segment for October 1974 to September 1993).



ACID = Anderson-Cottonwood Irrigation District.

Figure O.3-2. Spawning WUA curves for Winter-Run Chinook Salmon in the Sacramento River, Segments 4 to 6.



ACID = Anderson-Cottonwood Irrigation District.

Figure O.3-3. Spawning WUA Curves for Fall-Run Chinook Salmon in the Sacramento River, Segments 4 to 6. These Curves were used to Quantify Spring-run Spawning WUA, as Discussed in the Text.



ACID = Anderson-Cottonwood Irrigation District.

Figure O.3-4. Spawning WUA Curves for Late Fall–Run Chinook Salmon in the Sacramento River, Segments 4 to 6.



ACID = Anderson-Cottonwood Irrigation District.

Figure O.3-5. Spawning WUA curves for Steelhead in the Sacramento River, Segments 4 to  $6<sub>1</sub>$ 

To evaluate the relative importance of results from the river three segments (four segments for fall-run) for each of the salmon races, the typical spawning distributions of the races with respect to the segments (Table O.3-2) were estimated from the aerial redd surveys conducted by California Department of Fish and Wildlife during 2006 through 2021 (CDFW unpublished data). All races other than fall-run primarily spawn upstream of the Battle Creek confluence, and most fall-run spawning occurs upstream of the RBDD. Little is known about steelhead spawning locations in the Sacramento River, although it was assumed for this analysis that, because of constraints on water temperature and other habitat features, individuals spawn between Keswick Dam and RBDD, where nearly all Chinook salmon spawn (Table O.3-2). For the salmon races, the mean WUA results for the segments were weighted using the percentage in Table O.3-2 to compute total mean spawning WUAs. The total mean spawning WUA results, therefore, were computed to account for both the temporal (Table O.3-1) and spatial distributions (Table O.3-2) of spawning. WUA curves for steelhead were available only for Segments 4, 5, and 6 (Figure O.3-5). The steelhead spawning distribution among the three segments is uncertain, so the WUA results for the three segments were weighted equally in computing the total mean steelhead spawning WUA. Differences in the mean spawning WUA under the NAA and the seven management alternatives were examined for the months of the spawning periods of each race or species under each water year type and all water year types combined.

**Segment No. Description River Miles Winter-Run Spring-Run Fall-Run Late Fall– Run** 6 Keswick to ACID 302-298.5 35.6% 5.9% 17.4% 62.0% 5 ACID to Cow Creek 298.5-280 63.0% 72.1% 32.9% 19.8% 4 Cow Creek to Battle Creek 280-271 0.4% 6.7% 14.2% 8.7% 3 Battle Creek to RBDD 271-243 0.2% 3.6% 18.1% 3.7% 2 Downstream of RBDD — 0.8% 11.7% 17.4% 5.8%

Table O.3-2. Distributions of Spawning Redds among WUA River Segments as Percent of Total in the Sacramento River for Chinook Salmon Runs.

ACID = Anderson-Cottonwood Irrigation District RBDD = Red Bluff Diversion Dam

# *O.3.2.1.2 Sacramento River Rearing WUA*

The rearing habitat WUA curves used for Chinook salmon rearing habitat in the Sacramento River were obtained from a U.S. Fish and Wildlife Service report (U.S. Fish and Wildlife Service 2005b). As noted above for spawning habitat, WUA is computed as the surface area of physical habitat available weighted by its suitability. Modeling assumptions used to derive rearing WUA curves include that the suitability of physical habitat for salmon and steelhead rearing is largely a function of water depth, flow velocity, and the availability of cover. The race- or species-specific suitability of the habitat with respect to these variables is determined from field observations and measurements of habitat use by the fish, which is used to develop HSC for each race or species. Hydraulic modeling (using PHABSIM in the U.S. Fish and Wildlife Service 2005b study) is then used to estimate the amount of rearing habitat available for different HSC levels at different river flows, and the results are used to develop rearing habitat WUA curves and tables (Bovee et al. 1998). These curves and tables are used to look up the amount of rearing WUA available at different flows.

U.S. Fish and Wildlife Service (2005b) provides WUA curves and tables for rearing winter-run, fall-run, and late fall–run Chinook salmon for three segments of the Sacramento River encompassing the reach from Keswick Dam to Battle Creek (Figure O.3-1). Separate curves were developed for fry and juveniles, with fry defined as fish less than 60 millimeters and juveniles defined as greater than 60 millimeters. No WUA curves were developed for spring-run Chinook salmon or steelhead, but as discussed below in Section O.3.2.2, *Assumptions / Uncertainty*, the fall-run curves were used to quantify spring-run rearing habitat and the late fall– run curves were used for steelhead. Although fall-run rearing WUA curves were used as surrogates for spring-run rearing, CalSim 3 flows for the months of spring-run rearing, not those of fall-run rearing, were used to compute the spring-run WUA results. This caveat applies as well to the use of the late fall–run rearing WUA curves to compute steelhead rearing WUA results. Figure O.3-6 through Figure O.3-11 show the flow versus rearing WUA results for fry and juvenile winter-run, fall-run, and late fall–run Chinook salmon in the three river segments (Segment  $6 =$  Keswick to ACID Dam, Segment  $5 =$  ACID Dam to Cow Creek, and Segment  $4 =$ Cow Creek to Battle Creek) as provided in U.S. Fish and Wildlife Service (2005b). Note that for

Segment 6, separate WUA curves were developed for periods when the ACID Dam boards are installed (April through October) and for when the boards are out because installation of the boards affects water depths and flow velocities for some of the sampling transects used to develop the curves. All rearing WUA analyses were limited to juveniles less than a year old.

As previously noted, several tributaries enter the Sacramento River between Keswick Dam and Battle Creek, resulting in differences in flow among the river segments. For the U.S. Fish and Wildlife Service studies, flows were measured directly at the sampling transects and were estimated as the sum of Keswick flow releases and tributary gauge readings upstream of the transects. To estimate rearing WUA for this analysis, the segment flows were estimated using Sacramento River CalSim 3 flows at Keswick Dam and the confluences at Clear Creek and Battle Creek for Segments 6, 5, and 4, respectively. Keswick Dam flows were also used for Segment 5 upstream of the Clear Creek confluence. For Segment 6, the WUA curves for the months when the ACID Dam boards are installed (April through October) were used with the flows for those months and the WUA curves for the months when the ACID Dam boards are out were used with the flows for the rest of the year. Differences in the mean rearing WUA under conditions and alternatives were examined for the months of the fry and juvenile rearing periods for each race or species under each water year type and all water year types combined.

It should be noted that many winter-run fry begin moving downstream shortly after emerging and a majority of the fry may rear primarily downstream of the RBDD (Martin 2001). This may also be true for fry of the other salmon runs and steelhead. Unfortunately, no rearing WUA studies have been conducted for the Sacramento River downstream of the RBDD. Because of uncertainties and variability in the distribution of fry and juvenile rearing with respect to the three river segments for which rearing WUA curves were developed, results from the three segments were weighted equally in computing the total mean WUAs.

Mean fry rearing WUAs from each of the river segments were determined for the principal months of rearing for winter-run, spring-run, and steelhead fry (Table O.3-3) under each water year type and all water year types combined. Mean rearing WUA for all months combined was computed by weighting the monthly average results by monthly weighting factors (Table O.3-3). The weighting factors for winter-run, spring-run, and steelhead were estimated from results of the rotary screw trap monitoring at RBDD provided in the USFWS Sac PAS online database. The primary months of fry rearing for fall-run and late fall-run were obtained in consultations with NMFS for the Department of Water Resources California WaterFix Project. The months of the rearing period for these two races were weighted equally.



Table O.3-3. Monthly Weighting Factors for Fry Rearing of Sacramento River Winter-run, Spring-run, Fall-run, Late Fall-run and Steelhead.

The beginning of the juvenile (length >60 millimeters) rearing period was difficult to derive from field study data because of a high level of temporal overlap with the end of the fry rearing period. Therefore, the juvenile period was assumed to begin a fixed period after the start of the fry period. Fry upstream of RBDD have a growth rate of about 0.33 millimeters per day (Healey 1991) and the initial length of fry at emergence is about 40 millimeters (McMichael et al. 2005; Geist et al. 2006), so the juvenile period was determined to begin two months after the start of the fry period: October for winter-run, January for spring-run, February for fall-run, May for late fall-run, and June for steelhead. Young of year juveniles largely move downstream below RBDD by January for winter-run, by May for spring-run, and by September for steelhead (see Figures 2, 27, and 34, respectively, in LTO Appendix AB-C, Species Spatial-Temporal Domains). Therefore, juvenile rearing WUA was computed for October through January for winter-run, January through May for spring-run, and June through September for steelhead. The juvenile rearing periods upstream of RBDD for fall-run and late fall-run were assumed to be similar in duration to those of spring-run. No monthly weighting factors were used for these periods because monthly variations in abundance of the juveniles is highly uncertain. Mean fry and juvenile rearing WUA under the four phases of Alternative 2 and four management alternatives were examined for the months of the rearing periods of each race or species under each water year type and all water year types combined.



ACID = Anderson-Cottonwood Irrigation District.

Figure O.3-6. Rearing WUA Curves for Winter-Run Chinook Salmon Fry in the Sacramento River, Segments 4 to 6.



ACID = Anderson-Cottonwood Irrigation District.

Figure O.3-7. Rearing WUA Curves for Winter-Run Chinook Salmon Juveniles in the Sacramento River, Segments 4 to 6.



ACID = Anderson-Cottonwood Irrigation District.

Figure O.3-8. Rearing WUA Curves for Fall-Run Chinook Salmon Fry in the Sacramento River, Segments 4 to 6. These Curves were used to Quantify Spring-run Fry Rearing WUA, as Discussed in the Text.



ACID = Anderson-Cottonwood Irrigation District.

Figure O.3-9. Rearing WUA Curves for Fall-Run Chinook Salmon Juveniles in the Sacramento River, Segments 4 to 6. These Curves were used to Quantify Spring-run Juvenile WUA, as Discussed in the Text.



ACID = Anderson-Cottonwood Irrigation District.

Figure O.3-10. Rearing WUA Curves for Late Fall–Run Chinook Salmon Fry in the Sacramento River, Segments 4 to 6. These Curves were used to Quantify Steelhead Fry Rearing WUA, as Discussed in the Text.



ACID = Anderson-Cottonwood Irrigation District.

Figure O.3-11. Rearing WUA Curves for Late Fall–Run Chinook Salmon Juveniles in the Sacramento River, Segments 4 to 6. These Curves were used to Quantify Steelhead Juvenile Rearing WUA, as Discussed in the Text.

A potential limitation of all the WUA curves presented above, as of all such habitat-based studies, is that they assume the channel characteristics of the river during the time of field data collection by USFWS (1995–1999), such as proportions of mesohabitat types, have remained in dynamic equilibrium to the present time and would continue to do so through the life of the Project. If the channel characteristics substantially change, the shape of the curves may no longer be applicable. A further limitation of the rearing WUA curves is that they were developed for the Sacramento River upstream of Battle Creek, but all races of Chinook salmon and steelhead spend time rearing downstream of this part of the river.

# **O.3.2.2 Assumptions / Uncertainty**

This section includes two subsections. The first subsection provides a list of some important uncertainties and assumptions of the WUA analyses used for this analysis. The second subsection provides a more general discussion of the validity of WUA analysis, responding to concerns that have been raised in the scientific literature.

#### *O.3.2.2.1 Important Uncertainties and Assumptions of the WUA Analyses Conducted for the Effects Analyses*

The CalSim 3 operations model used to estimate spawning and rearing WUA under the phases and the alternatives employs a monthly timestep. Therefore, the WUA results should be treated as monthly averages. Monthly average WUA results faithfully represent the average conditions affecting the fish. Therefore, using monthly averages to compare WUA results is acceptable for showing differences in the effects of the different flow regimes under phases and alternatives conditions. Weighting by the weighting factors in Table O.3-1 and Table O.3-3 ensures that the comparisons account for differences in the amount of spawning occurring in each month, improving the validity of the results.

As noted previously, fall-run Chinook salmon WUA curves were used to model Sacramento River spring-run habitat in the analysis. This substitution follows previous practice. For instance, two models that currently produce spawning WUA outputs for spring-run Chinook salmon, SALMOD and Sacramento River Ecological Flows Tool (SacEFT), derive the spring-run WUA results using the fall-run Chinook salmon spawning WUA curves as surrogates (Bartholow 2004; ESSA Technologies 2011). Mark Gard, who led the U.S. Fish and Wildlife Service studies that produced the Sacramento River WUA curves, has endorsed this practice (Gard pers. comm.). This practice introduces additional uncertainty to the spring-run Chinook salmon results.

As described previously, the habitat suitability criteria used to develop the steelhead WUA curve for Sacramento River spawning were obtained from investigations of steelhead redds in the American River (U.S. Fish and Wildlife Service 2003b) because few steelhead redds were observed in the Sacramento River and the steelhead redds could not be distinguished from those of resident rainbow trout. The validity of this substitution could not be tested and is uncertain (U.S. Fish and Wildlife Service 2003a).

Rearing WUA curves were developed for the Sacramento River only for reaches upstream of Battle Creek, but all races of Chinook salmon and steelhead spend time rearing downstream of this part of the river. This limitation creates uncertainty regarding effects of the phases and alternatives on rearing habitat in the Sacramento River downstream of Battle Creek.

As previously discussed, no spring-run Chinook salmon or steelhead rearing WUA curves were developed in the U.S. Fish and Wildlife Service studies. Following previous practice, the fall-run and late fall–run Chinook salmon rearing WUA curves were used as surrogates in this analysis to model rearing habitat for spring-run and steelhead, respectively. Mark Gard, who led the U.S. Fish and Wildlife Service studies that produced the Sacramento River WUA curves, has endorsed this practice for both spring-run Chinook salmon and steelhead (Gard pers. comm.). The use of these substitutions has previously been adopted for the SacEFT model (ESSA Technologies 2011; Robinson pers. comm.). It should be noted that this practice introduces additional uncertainty to the spring-run and steelhead results.

The suitability of physical habitat for salmon and steelhead spawning is largely a function of substrate particle size, water depth, and flow velocity. Other unmeasured factors (e.g., flow vortices, water quality, etc.) could influence habitat suitability, contributing to uncertainty in the results.

The suitability of physical habitat for salmon and steelhead fry and juvenile rearing is largely a function of availability of cover, water depth, and flow velocity. Other unmeasured factors (e.g., flow vortices, complex feeding behaviors, water quality, etc.) could influence habitat suitability, contributing to uncertainty in the results.

The output of the WUA analysis, Weighted Usable Area, is an index of habitat suitability, not an absolute measure of habitat surface area. In the literature, Weighted Usable Area is often expressed as square feet, square meters, or acres for a given linear distance of stream, which is misleading and can result in unsupported conclusions (Payne 2003; Railsback 2016; Reiser and Hilgert 2018).

Both spawning and rearing WUA analyses assume that the channel characteristics of the river, such as proportions of mesohabitat types, during the time of field data collection by U.S. Fish and Wildlife Service (1995–1999) have remained in dynamic equilibrium to the present time and will continue to do so through the life of the Project. If the channel characteristics substantially changed, the shape of the curves might no longer be applicable.

#### *O.3.2.2.2 Discussion Regarding Validity of Weighted Usable Area Analysis*

WUA analysis is among the most widely used and recognized analytical tools for assessing effects of flow on fish populations (Reiser and Hilgert 2018). Procedures for quantifying WUA were developed and standardized by USFWS in the 1970s and they have since been widely adopted by researchers (e.g., Bourgeois et al. 1996; Beecher et al. 2010; Railsback 2016; Naman et al. 2020). However, WUA analysis has received some criticism from instream flow analysis practitioners, especially in recent years. Many conclusions in this analysis regarding effects on fish of changes in flow resulting from operations are based on WUA analyses. Therefore, it is important to understand and evaluate the criticisms of WUA analysis and consider any potential limitations for assessing flow-related effects.

Two frequent criticisms of the WUA analysis that are most potentially relevant with regard to the results and conclusions of the analysis are: (1) WUA analysis fails to directly evaluate many factors that are known to be important to fish population production, including water quality (especially temperature), predation, competition, and food supply (Beecher et al. 2010; Railsback 2016; Naman et al. 2019, 2020), and (2) the models employed to develop the WUA curves (especially PHABSIM) are antiquated, the field observations and measurements used to run the models are not sufficiently fine-grained to capture important highly localized factors, and the models do not adequately capture many dynamic properties of fish habitat use (Railsback 2016; Reiser and Hilgert 2018).

Regarding the first criticism, PHABSIM and the WUA curves they produce were never meant to address all factors affecting fish populations. As noted in a recent paper rebutting many of the criticisms of PHABSIM (Stalnaker et al. 2017): "PHABSIM is a component of instream flow incremental methodology (IFIM), which is a multifaceted decision support system that looks at riverine ecology for the purpose of making water management decisions." The IFIM uses a suite of evaluation tools (including PHABSIM) and investigates water quality factors and other factors that affect fish in addition to the hydraulic-related habitat conditions analyzed using PHABSIM or other hydraulic habitat models (Beecher 2017). These methods typically include evaluation tools for assessing effects of water temperatures, redd dewatering, adult migration passage,

emigrating juvenile salmonid survival, water diversion entrainment, and other factors. Analysis methods other than PHABSIM are used to evaluate the other factors, which may or may not be affected by flow. Conclusions regarding effects of the Project on a species are based on evaluations of the results for all the factors analyzed.

The second criticism is more specific to the modeling tools used for WUA analyses. Many of the limitations of PHABSIM cited by critics are acknowledged by its defenders (Beecher 2017; Stalnaker et al. 2017; Reiser and Hilgert 2018). Some of the cited shortcomings are common to any model that attempts to simulate complex ecological systems. Others reflect that PHABSIM is antiquated; newer, more powerful procedures have been incorporated into newer models. In fact, many studies have replaced PHABSIM with more powerful tools in recent years, including the RIVER2D hydraulic and habitat model that was used by USFWS to develop the Sacramento River spawning WUA curves used for the [Project] WUA analyses (U.S. Fish and Wildlife Service 2006). The field data used for the hydraulic/habitat modeling have also been refined and improved. For instance, improvements have been made in the flow velocity data used to represent the full range of flow velocity conditions affecting drift-feeding juvenile salmonids (Naman et al. 2019). The U.S. Fish and Wildlife Service studies of Sacramento River rearing WUA include such a modification to represent flow velocities (U.S. Fish and Wildlife Service 2005a). In addition, improvements have been developed to include a broader range of factors in the modeling, including some of those mentioned in the previous paragraph. One of these includes modeling of bioenergetic factors (Naman et al. 2020). Such methods are promising, but they are not currently available for use in analyzing flow effects on fish populations in the Sacramento River system.

Some shortcomings of WUA analysis are more difficult to remedy. For instance, competition within a cohort of juvenile salmonids may affect habitat use such that dominant fish exclude subdominants from optimal habitat locations, resulting in the highest densities of fish occupying sub-optimal habitat (Beecher et al. 2010; Beecher 2017). Some such biases are inevitable in any effort to model fish populations, but improvements in sampling and modeling techniques can be expected to lead to more accurate models in the future. PHABSIM and similar models, despite their shortcomings, continue to be among the most used and useful analytical tools for assessing instream-flow-related issues (Reiser and Hilgert 2018).

# **O.3.2.3 Code and Data Repository**

Data for this analysis is available from Reclamation upon request.

# **O.3.3 Results**

The following results provide the estimates of spawning and rearing WUA for winter-run, spring-run, fall-run, and late fall-run Chinook salmon and steelhead. For each race and species, the spawning and rearing WUA results are provided separately, with tables and figures for the BA and EIS modeled scenarios included in each section.

### **O.3.3.1 Winter-run Chinook Salmon**

#### *O.3.3.1.1 Spawning Weighted Usable Area*

Table O.3-4 and Table O.3-5 provide the spawning WUA results for Sacramento River winterrun Chinook salmon under the BA modeled scenarios and EIS modeled scenarios, respectively. The results are the means for all years analyzed, weighted by average proportion of redds counted per month (Table O.3-1) and per river segment (Table O.3-2). The table for the EIS modeled scenarios includes the percent differences between the results of the NAA and the alternatives (Table O.3-5).

The results for both the BA and EIS modeled scenarios show modest and inconsistent variation in mean spawning WUA with water year type for EXP1 and EXP3, but under the NAA and all BA and EIS modeled scenarios for the alternatives, the variation in mean spawning WUA among water year types is consistent, with the highest WUA under critically dry water years and lowest in above normal water years. For the EIS modeled scenarios, the largest difference between the NAA and the scenarios is a 5.6% increase for Alt 3 in above normal water years (Table O.3-5). The largest reduction is 1.4% for Alt 1 in above normal water years.

Table O.3-4. Expected WUA for Winter-run Chinook Spawning in the Sacramento River from Keswick Dam to the Battle Creek Confluence for EXP1, EXP3, NAA, and four phases of Alternative 2



Table O.3-5. Expected WUA for Winter-run Chinook Spawning in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1-4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA





Figure O.3-12 and Figure O.3-13 show the full variation in estimated spawning WUA for winterrun under the BA and EIS modeled scenarios, respectively. The upper and lower limits or the range in WUA values are determined by the ranges of the spawning WUA curves from which they are estimated (Figure O.3-2). The estimated spawning WUA values under the BA and EIS modeled scenarios are similar for May, June, and August, but the values are lower and more variable for July. The CalSim 3 flows are substantially higher in July than in the other months, which produces the lower and more variable spawning WUA results.



Figure O.3-12. Expected WUA for Winter-run Chinook Salmon Spawning in the Sacramento River Segments 6+5 for EXP1, EXP3, NAA, and four phases of Alternative 2 by Month



Figure O.3-13. Expected WUA for Winter-run Chinook Salmon Spawning in the Sacramento River Segments 6+5 for the NAA and Alternatives 1-4 by Month

# *O.3.3.1.2 Rearing Weighted Usable Area*

Table O.3-6 through Table O.3-9 provide the rearing WUA results for fry and juveniles of Sacramento River winter-run under the BA modeled scenarios and EIS modeled scenarios, respectively. The results are the means for all years analyzed, weighted by estimated monthly average abundance as discussed above in the methods section. The tables for the EIS modeled scenarios include the percent differences between the results of the NAA and the alternatives (Table O.3-7 and Table O.3-9).

The results for both the BA and EIS modeled scenarios show modest and inconsistent variation in mean fry rearing WUA with water year type for EXP 1 and EXP 3, but under the NAA and both the BA and EIS modeled scenarios for the alternatives, the variation in mean rearing WUA among water year types is generally consistent, with the highest WUA under dry or critically dry water years and lowest in above normal or wet water years (Table O.3-6 and Table O.3-7). However, the variation among water year types is small. For the EIS modeled scenarios, the largest difference between the NAA and the scenarios is a 3.2% increase for Alt 1 in above normal water years (Table O.3-7). The largest reduction is 1.5% for Alt 1 in critical water years.

Table O.3-6. Expected WUA for Winter-run Chinook Fry Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for EXP1, EXP3, the NAA, and four phases of Alternative 2



Table O.3-7. Expected WUA for Winter-run Chinook Fry Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1-4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA



Figure O.3-14 and Figure O.3-15 show the full variation in estimated fry rearing WUA for winter-run under the BA and EIS modeled scenarios, respectively. The upper and lower limits of the range in WUA values are determined by the ranges of the fry rearing WUA curves from which they are estimated (Figure O.3-6). The estimated fry rearing WUA values under the BA and EIS modeled scenarios are similar for the five months of winter-run fry rearing, August through December. However, extreme WUA results, particularly results with higher WUA values, are much more prevalent for December and somewhat more prevalent for November, presumably because of more frequent high flows. This result is due to the increasing rearing WUA values for higher flows in the winter-run fry rearing WUA curve (Figure O.3-7). Winterrun is the only Sacramento River salmonid race or species that shows this pattern.



Figure O.3-14. Expected WUA for Winter-run Chinook Salmon Fry Rearing in the Sacramento River Segments 4-6 for EXP1, EXP3, NAA, and four phases of Alternative 2 by Month



Figure O.3-15. Expected WUA for Winter-run Chinook Salmon Fry Rearing in the Sacramento River Segments 4-6 the NAA and Alternatives 1-4 by Month

The results for both the BA and EIS modeled scenarios show little variation in mean winter-run juvenile rearing WUA with water year type (Table O.3-8 and Table O.3-9). For the EIS modeled scenarios, the largest difference between the NAA and the scenarios is a 0.7% increase for Alt 4 in critical water years (Table O.3-9). The largest reduction is 0.5% for Alt2 Without TUCP Systemwide VA in below normal water years.

Table O.3-8. Expected WUA for Winter-run Chinook Juvenile Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for EXP1, EXP3, the NAA, and four phases of Alternative 2.



Table O.3-9. Expected WUA for Winter-run Chinook Juvenile Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1- 4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA



Figure O.3-16 and Figure O.3-17 show the full variation in estimated juvenile rearing WUA for winter-run under the BA and EIS modeled scenarios, respectively. The upper and lower limits of the range in WUA values are determined by the ranges of the juvenile rearing WUA curves from which they are estimated (Figure O.3-7). The estimated juvenile rearing WUA values under the BA and EIS modeled scenarios are similar for the five months of winter-run fry rearing, August through December. However, as reported for the fry rearing results (Figure O.3-14 and Figure O.3-15), extreme WUA results, particularly results with higher WUA values, are much more prevalent for November and December, presumably because of more frequent high flows. This result is due to the increasing rearing WUA values for higher flows in the winter-run juvenile rearing WUA curve (Figure O.3-7). As noted for the rearing WUA curve, winter-run is the only Sacramento River salmonid race or species that shows this pattern.



Figure O.3-16. Expected WUA for Winter-run Chinook Salmon Juvenile Rearing in the Sacramento River Segments 4-6 for EXP1, EXP3, NAA, and four phases of Alternative 2 by Month



Figure O.3-17. Expected WUA for Winter-run Chinook Salmon Juvenile Rearing in the Sacramento River Segments 4-6 for the NAA and Alternatives 1-4 by Month

# **O.3.3.2 Spring-run Chinook Salmon**

#### *O.3.3.2.1 Spawning Weighted Usable Area*

Table O.3-10 and Table O.3-11 provide the spawning WUA results for Central Valley spring-run Chinook salmon under the BA modeled scenarios and EIS modeled scenarios, respectively. The results are the means for all years analyzed, weighted by their expected distributions among months (Table O.3-1) and river segments (Table O.3-2). The table for the EIS modeled scenarios includes the percent differences between the results of the NAA and the alternatives (Table O.3-11).

The results for both the BA and EIS modeled scenarios show modest and inconsistent variation in mean spawning WUA with water year type for EXP 1 and EXP 3, but under the NAA and the BA and EIS modeled scenarios for alternatives, the variation in mean spawning WUA among water year types is generally consistent, with the highest WUA under critically dry or dry water years and lowest in above normal or wet water years. For the EIS modeled scenarios, the largest difference between the NAA and the scenarios is a 15.5% increase for Alt 1 in above normal water years (Table O.3-7). The largest reduction is 2.7% for Alt 1 in critical water years.

Table O.3-10. Expected WUA for Spring-run Chinook Spawning in the Sacramento River from Keswick Dam to the Battle Creek Confluence for EXP1, EXP3, the NAA, and four phases of Alternative 2.



Table O.3-11. Expected WUA for Spring-run Chinook Spawning in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1-4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA.





Figure O.3-18 and Figure O.3-19 show the full variation in estimated spawning WUA for springrun under the BA and EIS modeled scenarios, respectively. The upper and lower limits of the range in WUA values are determined by the ranges of the spawning WUA curves from which they are estimated (Figure O.3-3). The estimated spawning WUA values under the BA and EIS modeled scenarios are similar for September and October but are lower for August. The CalSim 3 flows are generally higher in August than in the other two months, which could result in lower spawning WUA results. The fall-run spawning WUA curves (Figure O.3-3), which were used to estimate spring-run spawning WUA, peak at relatively low flows (3,000 to 6,000 cfs).



Figure O.3-18. Expected WUA for Spring-run Chinook Salmon Spawning in the Sacramento River Segments 6+5 for EXP1, EXP3, NAA, and four phases of Alternative 2 by Month



Figure O.3-19. Expected WUA for Spring-run Chinook Salmon Spawning in the Sacramento River Segments 6+5 the NAA and Alternatives 1-4 by Month

# *O.3.3.2.2 Rearing Weighted Usable Area*

Table O.3-12 through Table O.3-13 provide the rearing WUA results for fry and juveniles of Sacramento River spring-run under the BA modeled scenarios and EIS modeled scenarios. The results are the means for all years analyzed, weighted by estimated monthly average abundance as discussed above in the methods section. The tables for the EIS modeled scenarios include the percent differences between the results of the NAA and the alternatives (Table O.3-7 and Table O.3-9).

The results for both the BA and EIS modeled scenarios show consistent variation in mean springrun fry rearing WUA with water year type for EXP 1, EXP 3, the NAA and all BA and EIS modeled scenarios for alternatives, with the highest WUA under critically dry water years and lowest wet water years (Table O.3-12 and Table O.3-13). EXP 1 is an exception, having the lowest WUA value for above normal water years. For the EIS modeled scenarios, the fry rearing WUA results were generally lower under the EIS modeled scenarios than under the NAA. The largest reductions between the NAA and the scenarios are 2.6% reductions for Alt2 Without TUCP Delta VA and Alt2 Without TUCP Systemwide VA under critical water years (Table O.3-13).

Table O.3-12. Expected WUA for Spring-run Chinook Fry Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for EXP1, EXP3, the NAA, and four phases of Alternative 2.



Table O.3-13. Expected WUA for Spring-run Chinook Fry Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1-4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA.



Figure O.3-20 and Figure O.3-21 show the full variation in estimated fry rearing WUA for spring-run under the BA and EIS modeled scenarios, respectively. The upper and lower limits of the range in WUA values are determined by the ranges of the fall-run fry rearing WUA curves from which they are estimated (Figure O.3-8). The estimated fry rearing WUA values under the BA and EIS modeled scenarios are similar for the principal months of spring-run fry rearing, December through March, but they are generally lower for November and April (Figure O.3-16 and Figure O.3-17). Under the EIS modeled alternatives, the December through March period, results show little variation in the third quartile of the WUA distributions, except for Alternative 1. This occurs because the fry rearing WUA curves peak at the lowest flows encountered in the river (Figure O.3-8) and Keswick flow releases are frequently low during these months in critically dry water year types when Shasta Dam is operated to rebuild storage volume. The second quartile is much wider because a greater range of flows affect the fry rearing WUA values below the median. The occurrence of some higher rearing WUA values for April results from the installation of the ACID dam for that month, which results in higher WUA values in Segment 6 of the river (Figure O.3-8).



Figure O.3-20. Expected WUA for Spring-run Chinook Salmon Fry Rearing in the Sacramento River Segments 4-6 for EXP1, EXP3, NAA, and four phases of Alternative 2 by Month



Figure O.3-21. Expected WUA for Spring-run Chinook Salmon Juvenile Rearing in the Sacramento River Segments 4-6 for the NAA and Alternatives 1-4 by Month



The results for both the BA and EIS modeled scenarios show consistent variation in mean springrun juvenile rearing WUA with water year type for EXP 1, EXP 3, the NAA and all BA and EIS modeled scenarios for alternatives, with the highest WUA under critically dry water years and lowest under wet water years (Table O.3-12 and Table O.3-13). For the EIS modeled scenarios, the fry rearing WUA results are generally modestly lower under the EIS modeled scenarios than under the NAA. The largest reduction between the NAA and the scenarios is 1.7% for Alt 3 under critical water years (Table O.3-15).

Table O.3-14. Expected WUA for Spring-run Chinook Juvenile Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for EXP1, EXP3, the NAA, and four phases of Alternative 2.



Table O.3-15. Expected WUA for Spring-run Chinook Juvenile Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1- 4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA.



Figure O.3-22 and Figure O.3-23 show the full variation in estimated juvenile rearing WUA for spring-run under the BA and EIS modeled scenarios, respectively. The upper and lower limits of the range in WUA values are determined by the ranges of the fall-run juvenile rearing WUA curves from which they are estimated (Figure O.3-9). The estimated juvenile rearing WUA values under the BA and EIS modeled scenarios are similar for January through March, but they are generally lower for April and May (Figure O.3-22 and Figure O.3-23). For the January through March period, as described above for spring-run fry rearing WUA, the third quartiles of the WUA distributions have high values and reduced variation. This occurs because, as described for the fry rearing WUA curves, the juvenile rearing WUA curves peak at the lowest flows encountered in the river (Figure O.3-9) and low flows are frequent during January through March when Shasta Dam is operated to rebuild storage.



Figure O.3-22. Expected WUA for Spring-run Chinook Salmon Juvenile Rearing in the Sacramento River Segments 4-6 for EXP1, EXP3, NAA, and four phases of Alternative 2 by Month



Figure O.3-23. Expected WUA for Spring-run Chinook Salmon Juvenile Rearing in the Sacramento River Segments 4-6 for the NAA and Alternatives 1-4 by Month

# *O.3.3.2.3 Steelhead*

# *O.3.3.2.4 Spawning Weighted Usable Area*

Table O.3-16 and Table O.3-17 provide the spawning WUA results for steelhead under the BA modeled scenarios and EIS modeled scenarios, respectively. The results are the means for all years analyzed, weighted by their expected distributions among months (Table O.3-1). As noted earlier, the distribution of steelhead spawning among the river segments is unknown, so WUA results were weighted equally for the three upper river segments. The table for the EIS modeled scenarios includes the percent differences between the results of the NAA and the alternatives (Table O.3-17).

The results for both the BA and EIS modeled scenarios show strong and consistent reductions in the mean spawning WUA from the driest to the wettest water year types (Table O.3-16 and Table O.3-17). This pattern of variation holds for EXP 1, EXP 3, the NAA, and all BA and EIS modeled scenarios of the alternatives. For the EIS modeled scenarios, differences between the NAA and the scenarios are consistently small. The largest difference is a 2.0% reduction for Alt 1 in above normal water years (Table O.3-17). Most other differences are less than 1%.

Table O.3-16. Expected WUA for Steelhead Spawning in the Sacramento River from Keswick Dam to the Battle Creek Confluence for EXP1, EXP3, the NAA, and four phases of Alternative 2.



Table O.3-17. Expected WUA for Steelhead Spawning in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1-4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA.



Figure O.3-24 and Figure O.3-25 show the full variation in estimated spawning WUA for steelhead under the BA and EIS modeled scenarios, respectively. The upper and lower limits of the range in WUA values are determined by the ranges of the spawning WUA curves from which they are estimated (Figure O.3-5). The medians of the estimated spawning WUA values under the BA and EIS modeled scenarios are similar for January through March, but they are higher for April (Figure O.3-24 and Figure O.3-25). For all months, the results show little variation in the third quartile of the WUA distributions. This occurs because the steelhead spawning WUA curves peak at low flows (Figure O.3-5) and Keswick flow releases are frequently low during these months when Shasta Dam is operated to rebuild storage volume. The second quartile is much wider because a greater range of flows affect the spawning WUA values below the median. This is especially true for February because flow below Keswick Dam is most variable in this month. The higher spawning WUA values for April result from installation of the ACID dam in that month, which results in much higher WUA values in Segment 6 of the spawning WUA curve (Figure O.3-5).



Figure O.3-24. Expected WUA for Steelhead Spawning in the Sacramento River Segments 4-6 for EXP1, EXP3, NAA, and four phases of Alternative 2 by Month



Figure O.3-25. Expected WUA for Steelhead Spawning in the Sacramento River Segments 4-6 for the NAA and Alternatives 1-4 by Month

# *O.3.3.2.5 Rearing Weighted Usable Area*

Table O.3-18 through Table O.3-21 provide the rearing WUA results for fry and juveniles of Sacramento River steelhead under the BA modeled scenarios and EIS modeled scenarios, respectively. The results are the means for all years analyzed, weighted by estimated monthly average abundance as discussed above in the methods section. The tables for the EIS modeled scenarios include the percent differences between the results of the NAA and the alternatives (Table O.3-19 and Table O.3-21).

The results for both the BA and EIS modeled scenarios show inconsistent variation in mean steelhead fry rearing WUA with water year type for EXP 1 and EXP 3, but generally consistent, moderate reductions in rearing WUA from drier to wetter water years for the NAA and all BA and EIS modeled scenarios alternatives (Table O.3-18 and Table O.3-19). For the EIS modeled scenarios, the fry rearing WUA results are generally similar or slightly lower under the EIS modeled scenarios than under the NAA, except during critical water years for which rearing WUA was up to 5.8% higher under Alternative 2 With TUCP Without VA (Table O.3-19). Alt 1 was an exception, with essentially no change from the NAA in critical water years.

Table O.3-18. Expected WUA for Steelhead Fry Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for EXP1, EXP3, the NAA, and four phases of Alternative 2.



Table O.3-19. Expected WUA for Steelhead Fry Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1-4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA.



Figure O.3-26 and Figure O.3-27 show the full variation in estimated fry rearing WUA for spring-run under the BA and EIS modeled scenarios, respectively. The upper and lower limits of the range in WUA values are determined by the ranges of the late fall-run fry rearing WUA curves from which they are estimated (Figure O.3-10). The estimated fry rearing WUA values under the BA and EIS modeled scenarios are highly variable among the months, April through October, of the fry rearing period (Figure O.3-26 and Figure O.3-27). WUA values of the fry rearing WUA curve are lowest in the range of 6,000 cfs to 27,000 cfs (Figure O.3-10) and flows in this range are almost twice as frequent during May through August as in April, September, or October. The April, September, and October results show much more variability in the WUA values than the May through August results, reflecting greater variability in flows during those three months.



Figure O.3-26. Expected WUA for Steelhead Fry Rearing in the Sacramento River Segments 4-6 for EXP1, EXP3, NAA, and four phases of Alternative 2 by Month



# Figure O.3-27. Expected WUA for Steelhead Fry Rearing in the Sacramento River Segments 4-6 for the NAA and Alternatives 1-4 by Month

The results for both the BA and EIS modeled scenarios show consistent variation in mean steelhead juvenile rearing WUA with water year type for EXP 1, EXP 3, the NAA, and all BA and EIS modeled scenarios for alternatives, with the highest WUA under critically dry water years and lowest wet water years (Table O.3-20 and Table O.3-21). The only exception to this pattern of variation is for EXP1, which has lower rearing WUA under below normal years than in above normal or wet year water years. For the EIS modeled scenarios, the fry rearing WUA results were generally moderately lower under the EIS modeled scenarios than under the NAA, with a maximum reduction of 1.6% for Alt 3 in critical water years (Table O.3-21).

Table O.3-20. Expected WUA for Steelhead Juvenile Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for EXP1, EXP3, the NAA, and four phases of Alternative 2.



Table O.3-21. Expected WUA for Steelhead Juvenile Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1-4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA.



Figure O.3-28 and Figure O.3-29 show the full variation in estimated juvenile rearing WUA for steelhead under the BA and EIS modeled scenarios, respectively. The upper and lower limits of the range in WUA values are determined by the ranges of the juvenile rearing WUA curves for late fall-run from which they are estimated (Figure O.3-11). The analysis for steelhead juvenile includes all months because steelhead juveniles may reside in freshwater for a year or more before emigrating to the sea. The estimated juvenile rearing WUA values under the BA and EIS modeled scenarios are similar for December through March, and they are lower for May through August (Figure O.3-28 and Figure O.3-29). As discussed previously for spring-run rearing WUA and steelhead spawning WUA, the high WUA values with low variability in the third quartile for December through March result from the peaking of the WUA curve at the lowest flows (Figure O.3-11) and frequent low Keswick Dam flow releases during these months to build Shasta storage. The wide range of values in the second quartile of these months reflects the high frequency and variability of higher flows in these wet months. The low juvenile rearing WUA values during June through August result from relatively frequent flows above about 10,000 cfs during those months and the low WUA values in the juvenile rearing WUA curves at flows greater than about 10,000 cfs.



Figure O.3-28. Expected WUA for Steelhead Juvenile Rearing in the Sacramento River Segments 4-6 for EXP1, EXP3, NAA, and four phases of Alternative 2 by Month



Figure O.3-29. Expected WUA for Steelhead Juvenile Rearing in the Sacramento River Segments 4-6 for the NAA and Alternatives 1-4 by Month

### **O.3.3.3 Fall-run Chinook Salmon**

#### *O.3.3.3.1 Spawning Weighted Usable Area*

Table O.3-22 provides the spawning WUA results for Central Valley fall-run Chinook salmon under the BA modeled scenarios and EIS modeled scenarios, respectively. The results are the means for all years analyzed, weighted by their expected distributions among months (Table O.3-1) and river segments (Table O.3-2). The table for the EIS modeled scenarios includes the percent differences between the results of the NAA and the alternatives (Table O.3-17).

The results for the EIS modeled scenarios show modest and inconsistent variation in mean spawning WUA with water year type among the NAA, but under the alternatives, the variation in mean spawning WUA among water year types is consistent, with the highest WUA under critically dry water years and lowest in wet water years. For the EIS modeled scenarios, the largest difference between the NAA and the scenarios is a 4.2% increase for Alternative 1 in above normal water years (Table O.3-22). The largest reduction is 2.5% for Alternative 3 in below normal water years.

Table O.3-22. Expected WUA for Fall-run Chinook Spawning in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1-4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA.



Figure O.3-30 shows the full variation in estimated spawning WUA for fall-run Chinook salmon under the EIS modeled scenarios. The upper and lower limits of the range in WUA values are determined by the ranges of the spawning WUA curves from which they are estimated (Figure O.3-4). The medians of the estimated spawning WUA values under the EIS modeled scenarios are similar throughout the spawning period (Figure O.3-30). October and September have less variation in WUA values in comparison to November and December, which is due to more variable flows expected with winter storms during the latter period.



Figure O.3-30. Expected WUA for Fall-run Spawning in the Sacramento River Segments 4-6 for the NAA and Alternatives 1-4 by Month

#### *O.3.3.3.2 Rearing Weighted Usable Area*

Table O.3-23 and Table O.3-24 provide the rearing WUA results for fry and juveniles of Sacramento River fall-run under the EIS modeled scenarios, respectively. The results are the means for all years analyzed, weighted by estimated monthly average abundance as discussed above in the methods section. The tables include the percent differences between the results of the NAA and the alternatives (Table O.3-23 and Table O.3-24).

The results show consistent, moderate variation in mean fall-run fry rearing WUA with water year type among the modeled scenarios, with the highest WUA under critically dry water years and lowest in wet water years (Table O.3-23). The fry rearing WUA results were mostly lower under the alternatives than under the NAA. The largest reductions between the NAA and the scenarios is a 2.6% reductions for Alternative 1 under above normal water years (Table O.3-12).

As was the case for the fall-run fry rearing WUA results, the results for fall-run juvenile rearing WUA show consistent variation in mean WUA with water year type among the NAA and Alternatives 1-4, with the highest WUA under critically dry water years and lowest in wet water years (Table O.3-24). The amount of reduction in juvenile rearing WUA is less than that found

for fry rearing WUA (Table O.3-23). For the EIS modeled scenarios, most of the juvenile rearing WUA values are similar between the EIS modeled scenarios and the NAA. The largest difference is a 3.5% increase for Alternative 3 in critical water years and the largest reduction is 2.0% under Alternative 1 in critical water years (Table O.3-24).

Table O.3-23. Expected WUA for Fall-run Chinook Fry Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1-4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA.



Table O.3-24. Expected WUA for Fall-run Chinook Juvenile Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1- 4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA.





Figure O.3-31 and Figure O.3-32 show the full variation in estimated fry and juvenile rearing WUA for fall-run under the EIS modeled scenarios. The upper and lower limits of the range in WUA values are determined by the ranges of the fall-run fry and juvenile rearing WUA curves from which they are estimated (Figure O.3-8 and Figure O.3-9). The estimated fry rearing WUA values under the EIS modeled scenarios are highly variable among the months, December through March (Figure O.3-31). However, the median values for fry rearing are relatively similar across all months for the NAA and Alternatives 2 and 4, with some reductions under Alternative 1 and 3 in comparison to the NAA. During the juvenile rearing period, the median values are similar amongst Alternatives 1-4 and the NAA between February and March but show differences April through June (Figure O.3-32). In April, Alternative 3 has the highest median value and Alternative 2 without TUCP Systemwide VA has the lowest median value. In June, when flows tend to be lowest during the juvenile rearing period, Alternative 3 has the highest median value and Alternative 1 has the lowest median value.



Figure O.3-31. Expected WUA for Fall-run Fry Rearing in the Sacramento River Segments 4-6 for the NAA and Alternatives 1-4 by Month



Figure O.3-32. Expected WUA for Fall-run Juvenile Rearing in the Sacramento River Segments 4-6 for the NAA and Alternatives 1-4 by Month

# **O.3.3.4 Late Fall-run Chinook Salmon**

### *O.3.3.4.1 Spawning Weighted Usable Area*

Table O.3-25 provides the spawning WUA results for late fall-run Chinook salmon under the EIS modeled scenarios. The results are the means for all years analyzed, weighted by their expected distributions among months (Table O.3-1) and river segments (Table O.3-2). The table for the EIS modeled scenarios includes the percent differences between the results of the NAA and the alternatives (Table O.3-25).

The results for the EIS modeled scenarios show large and consistent variation in mean late fallrun spawning WUA with water year type for the NAA and Alternatives 1-4, with the highest WUA under critically dry water years and lowest in wet water years (Table O.3-25). Mean WUA values are lower under Alternatives 1-4 than the NAA for almost all water year types. The largest reduction was 7.4% for Alternative 1 in above normal water years.

Table O.3-25. Expected WUA for Late Fall-run Chinook Spawning in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1- 4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA.



Figure O.3-33 shows the full variation in estimated spawning WUA for late fall-run under the EIS modeled scenarios. The upper and lower limits of the range in WUA values are determined by the ranges of the spawning WUA curves from which they are estimated (Figure O.3-5). The median values of the estimated spawning WUA under the EIS modeled scenarios are similar throughout the spawning period of January through March (Figure O.3-33).



Figure O.3-33. Expected WUA for Late Fall-run Spawning in the Sacramento River Segments 4-6 for the NAA and Alternatives 1-4 by Month

# *O.3.3.4.2 Rearing Weighted Usable Area*

Table O.3-26 and Table O.3-27 provide the rearing WUA results for fry and juveniles of late fallrun under the EIS modeled scenarios. The results are the means for all years analyzed, weighted by estimated monthly average abundance as discussed above in the methods section. The tables for the EIS modeled scenarios include the percent differences between the results of the NAA and the alternatives.

The results for the EIS modeled scenarios show moderate, consistent variation in mean late fallrun fry rearing WUA with water year type among Alternatives 1-4 and the NAA, with the highest WUA under critically dry water years and lowest under wet water years (Table O.3-26). The fry rearing WUA results are mostly similar between Alternatives 1-4 and the NAA. The largest difference between the NAA and the scenarios is a 5.6% increase for Alternative 3 under critical water years and the largest reduction is 3.5% for Alternative 1 in critical water years (Table O.3-26).

The results for late fall-run juvenile rearing WUA for the EIS modeled scenarios show little variation in mean WUA with water year type among the NAA and Alternatives 1-4 (Table O.3-27). The juvenile rearing WUA values for Alternatives 1-4 were generally higher than those of the NAA. The largest increase was a 6.7% for Alternative 2 With TUCP Without VA in critical water years (Table O.3-27). The largest reduction was 2.1% under Alternative 1 in critical water years.

Table O.3-26. Expected WUA for Late Fall-run Chinook Fry Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1- 4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA.



Table O.3-27. Expected WUA for Late Fall-run Chinook Juvenile Rearing in the Sacramento River from Keswick Dam to the Battle Creek Confluence for the NAA and Alternatives 1-4. The Lower Panel Gives the Percent Differences of the Alternatives and the NAA.





Figure O.3-34 and Figure O.3-35 show the full variation in estimated fry and juvenile rearing WUA for late fall-run under the EIS modeled scenarios. The upper and lower limits of the range in WUA values are determined by the ranges of the late fall-run fry and juvenile rearing WUA curves from which they are estimated (Figure O.3-10 and Figure O.3-11). The estimated fry rearing WUA values under the EIS modeled scenarios are variable among the months, March through June (Figure O.3-34). The median values for fry rearing vary to a lesser degree, with the most variation in April amongst the NAA and Alternatives 1-4. Between the months of the fry rearing period, May and June have the lowest expected WUA median values. During the juvenile rearing period, the median WUA values steadily decline from March to the lowest median WUA values in July, and rise to the highest median WUA values in September (Figure O.3-35).



Figure O.3-34. Expected WUA for Late Fall-run Fry Rearing in the Sacramento River Segments 4-6 for the NAA and Alternatives 1-4 by Month



Figure O.3-35. Expected WUA for Late Fall-run Juvenile Rearing in the Sacramento River Segments 4-6 for the NAA and Alternatives 1-4 by Month

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#### **O.3.4.1 Personal Communications.**

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